



Survival of rock wrens (*Xenicus gilviventris*) using radio-tags, through an aerial 1080 pest control operation

Tristan E. Rawlence^{1*} and Kim N. Squire¹

¹Department of Conservation, Private Bag 5, Nelson

*Author for correspondence: (Email: trawlence@doc.govt.nz)

Published online: 24 October 2024

Abstract: We monitored 15 rock wrens *Xenicus gilviventris* wearing very high frequency radio-tags and/or colour bands following an aerial 1080 pest control operation in Kahurangi National Park. We found no evidence that rock wrens were susceptible to mortality from 1080, but we did note some welfare concerns for rock wrens from carrying radio-tags. Due to the high effort required to monitor rock wrens using radio-tags coupled with the low risk to rock wrens from pest control operations, we recommend using non-invasive transect counts to monitor rock wren populations annually at multiple sites.

Keywords: alpine bird, monitoring, mortality, poisoning, predator control, radio-tag

Introduction

Introduced exotic mammals are implicated in the decline and extinction of many native species in New Zealand (Holdaway 1999; Robertson et al. 2021). The scale of control operations for rats (*Rattus* spp.), house mouse (*Mus musculus*), mustelids (*Mustela* spp.), and brushtail possums (*Trichosurus vulpecula*) has been increasing steadily over the last 25 years. The main tool used to control these exotic mammals on mainland New Zealand is aerially-applied cereal baits laced with 1080 poison (Elliott & Kemp 2016; Nugent et al. 2019). Currently, predators are controlled in aerial operations over vast (> 300 000 ha) areas of indigenous forest and alpine grassland, with the total area controlled exceeding 1 000 000 ha in some years (Elliott & Kemp 2016; Department of Conservation 2022).

In general, the fledging rate of native birds can be increased by reducing predator numbers (Wright 2011). However, some native species are vulnerable to primary and/or secondary poisoning from 1080 operations (Veltman & Westbrooke 2011; Kemp et al. 2019). For these species, more detailed research is required to evaluate the population impact of the loss of some individuals against the benefits of reduced predator abundances (Powlesland et al. 1999; Powlesland et al. 2000; Kemp et al. 2018).

Rock wrens (*Xenicus gilviventris*; also known as, pīwauwau, mātuitui, and tuke) are small (14–20 g) endangered alpine passerines inhabiting Kā Tiritiri o te Moana / Southern Alps of Te Waipounamu / South Island, New Zealand. They are the only specialist alpine bird in New Zealand, being the only species that lives and breeds exclusively above the treeline. They are one of only two extant species of the New Zealand endemic Acanthisittidae wrens, of which there were once eight species in six genera, making them the (formerly) most diverse group of passerines in New Zealand (Mitchell et al. 2016).

Rats have been largely responsible for the extinction of the rest of the New Zealand wrens (Holdaway 1999; Galbreath & Brown 2004; Bell et al. 2016).

There was a long-held belief that rock wrens, and alpine ecosystems, were relatively safe from introduced predators (Buller 1873; Guthrie-Smith 1936). However, more recent work has demonstrated that seasonal impacts from introduced predators can be significant (Stocker et al. 2006; Michelsen-Heath & Gaze 2007; O'Donnell et al. 2017; Weston et al. 2018; Rawlence 2019; Carpenter et al. 2023). The effects of human-induced climate change likely extend the duration of seasonal impacts, as temperature-limited ship rats (*Rattus rattus*) can inhabit alpine ecosystems for longer periods (Christie et al. 2017).

Increased awareness that predators need controlling in alpine ecosystems has led to an adaptive management approach to pest control in alpine areas, with a focus on outcomes for native species (Weston et al. 2018; Rawlence, 2019). Rawlence (2019) demonstrated that 1080 was effective in reducing predator numbers in the alpine areas of Kahurangi National Park and that rock wrens had significantly higher reproductive success in areas that received 1080 for pest control than in those that did not. This previous research instigated the current study, as some rock wrens disappeared after a 1080 operation in 2014, indicating the need to quantify potential primary or secondary poisoning risk to rock wrens from 1080.

The aim of this study was to better quantify the risk to rock wrens of aerial 1080 operations by using a more robust monitoring technique. In this study we attached very high frequency (VHF) radio-tags and colour bands to rock wrens and monitored them through an aerial 1080 operation in Kahurangi National Park in 2019.

Study Site

This study was undertaken on the flanks of Mt Domett (41.06°S, 172.32°E) in Kahurangi National Park, where rock wrens are resident from the sub alpine scrub (1180 m a.s.l.) to the mountain tops (1645 m a.s.l.). The study site (240 ha) is comprised of granite geology which has been moulded by prehistoric glaciation and earthquakes and receives 4–6 m of annual rainfall. The study was undertaken between 12 November and 13 December 2019 during the laying and incubation phases of the rock wren breeding season (Heath 1989; Rawlence 2019).

Methods

Rock wrens were initially located using playback calls in suitable habitat. Rock wrens were then herded into a 2.6 × 6 m mist net (30 mm polyester, Avinet) by a team of five people. Captured rock wrens were fitted with a unique combination of three coloured split bands (Darvic, 2.8 mm internal diameter). Metal bands were not used due to potential thermal risk to birds when in torpor (McNab & Weston 2020). Rock wrens were then fitted with a small VHF radio-tag (Holohil BD-2NT) with an elastic leg loop harness (Rappole & Tipton 1991), with the modifications described in Streby et al. (2015) (Fig. 1). The total weight of the harness and radio-tag was 0.51 g, which is 3% of the mean female rock wren weight and 3.2% of the mean male rock wren weight ($n = 314$; Department of Conservation, unpubl. data), below the universally agreed 5% rule (Barron et al. 2010). To keep weight down, these radio-tags had an expected battery duration of only 20 days.

The study site was nested within the ‘Oparara 2019’ aerial 1080 operation; this covered 39 324 ha and delivered non-toxic, cinnamon lured cereal baits (6 g RS5) to the entire study site at 2 kg ha⁻¹ as a prefeed on 3 November 2019, followed by toxic, cinnamon lured cereal baits laced with 0.15% 1080 poison (6 g RS5) at 2 kg ha⁻¹ on 22 November 2019.

All rock wrens were radio-tracked (Lotek, Biotracker VHF) and sighted the day before the 1080 operation, and subsequently at daily intervals as the weather permitted. Dead birds were located and remains were sent to Manaaki Whenua–Landcare Research for 1080 residue testing with a minimum detectable level (MDL) of 0.001 µg g⁻¹. Recapture for radio-tag removal was attempted for all birds where practical.

Results

We attached radio-tags and colour bands to 14 rock wrens (seven male, seven female), of which 13 were still wearing their radio-tags on the day of the 1080 operation. We were also able to monitor one rock wren (male) from a previous study who was still colour banded, along with the bird that dropped its radio-tag (male), as they were the partners of radio-tagged rock wrens from this study. This gave us a total of 15 marked birds that we could monitor.

All radio-tagged and banded rock wrens were sighted alive each day for eight days after the 1080 operation. Between 1 December and 9 December 2019 there was a prolonged period of heavy rain, with a total of 282 mm recorded at a nearby rain station. This rainfall event prevented birds being re-sighted for nine days. From 10 December 2019 (18 days after the operation) we managed to relocate all radio-tagged and banded birds except for one bird, whose radio-tag had exceeded its battery life. Two of the radio-tagged rock wrens were found dead (both female). One was found dead on its nest, but no 1080 was detectable in the muscle tissue. A second bird was preyed upon, with insufficient remains to test for 1080. This bird was most likely preyed upon by a falcon (*Falco novaeseelandiae*), based on the presence of plucked feathers found with the radio-tag and the absence of any other remains (Fig. 2), which is consistent with falcon feeding behaviour (Fox 1977).

We had some welfare concerns for the rock wrens wearing radio-tags during this study, one of which we could mitigate. Two of the first four birds fitted with full length radio-tag aerials showed sign of entanglement from the aerial, after which we shortened the aerial from 135 mm to 85 mm. No further entanglements were detected, nor was there a noticeable decrease in signal strength. We also observed higher than expected nest abandonments in birds fitted with radio-tags. From the 15 birds which had either radio-tags or colour bands attached, we found seven nests with known outcomes. Of these nests, three continued as normal, two failed because one of the parent birds died, and two pairs abandoned the nest, with one of these pairs re-nesting only to also abandon their second attempt.

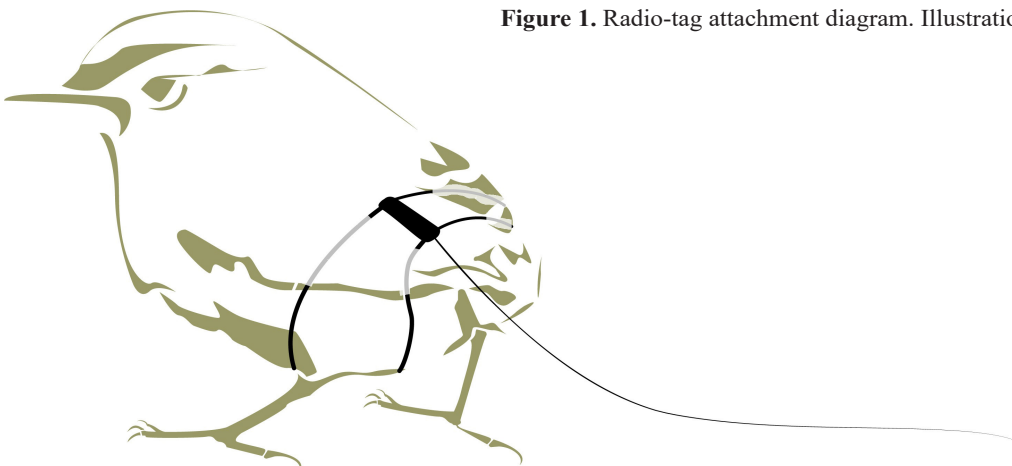


Figure 1. Radio-tag attachment diagram. Illustration by Anja McDonald.



Figure 2. Plucked feathers and radio-tag from the presumed falcon predated rock wren. Photo: author.

Discussion

The results of this study suggest that rock wrens are not susceptible to either primary or secondary poisoning in 1080 operations. All 15 rock wrens survived the operation. We would expect primary poisoning to have occurred within 8 days following the sowing of toxic baits, because 100 mm of rain (received after 9 days of monitoring) is the accepted level at which RS5 toxic baits are deemed non-toxic; this is due to the breakdown of the toxin upon contact with water (Bowen et al. 1995). Peak 1080 concentrations in invertebrates occur during the first 14 days following exposure; 1080 is still detectable in some invertebrates for a further 2 weeks, but at a low and diminishing level (Eason et al. 1993). Therefore, we would expect secondary poisoning to have occurred within 14 days following exposure.

After the 1080 risk period in this study had passed, we had one rock wren dead of unknown causes, one killed by a predator (probable New Zealand falcon), and another that we could not track due to radio-tag failure. Similar studies that retrieved dead passerines that later tested positive for 1080 recorded levels between $0.37\text{--}3.8\ \mu\text{g g}^{-1}$ (Powlesland et al. 1999; Powlesland et al. 2000). The single dead rock wren that we were able to assay did not test positive for 1080, which suggests 1080 did not enter the food chain of rock wrens during this study.

Veltman et al. (2014) suggested that 100 individuals need to be monitored, with no deaths attributed to 1080, for conservation managers to be confident of a $< 3\%$ mortality rate for any given species. The 15 radio-tagged and banded birds in this study are well short of the overall target, but in line with their suggestion of individual studies monitoring as

few as 15 individuals, which can later be pooled to achieve the 100 individual target.

Based on the findings of this study, we suggest that the disappearance of 22 rock wrens following a 1080 operation in 2014 (Rawlence 2019) was more likely to have been caused by an unseasonable snowstorm than 1080 poisoning. Furthermore, Weston et al. (2018) also noted nest abandonments and missing banded rock wrens after snowstorms during the breeding season. However, we cannot entirely rule out an interactive effect where rock wrens become trapped by heavy snow falls immediately after a 1080 operation, and are thus more likely to eat baits or poisoned invertebrates.

Monitoring rock wrens through 1080 operations using radio-tags is risky due to the welfare concerns observed during this study. Monitoring using radio-tags is also expensive due to the fickle nature of weather-dependent 1080 operations, the short 20-day battery life of radio-tags suitable for rock wrens, and the remote locations of most rock wren populations. In the 104 nests monitored by Rawlence (2019), abandonments only occurred due to snowstorms or poor trail camera installations at nests, never as a result of handling, which makes the abandonments observed in this study unusual, as no snow fell during the course of this study.

We believe that the welfare concerns observed during this study could be mitigated should the attachment of radio-tags be required in the future. We recommend having pre-made harnesses to reduce handling time, shortening aerials to

85 mm, and not attaching radio-tags during the nesting season, especially on gravid females. It would also be prudent to use rock wren populations without extensive cliffs in their home range to increase recapture chances for radio-tag removal.

Studies on small birds using radio-tags with a short battery life only provide a snapshot of how predator control operations impact a species, which is important if the risk of poisoning is high and carcass recovery is needed, but less important if poisoning risk is low, as observed in this study. The use of colour bands does provide a long-term method of monitoring, but is inadequate as a tool for monitoring mortality after 1080 operations, because carcasses are near impossible to find and missing birds can lead to inconclusive findings (Rawlence 2019).

Given the cost to rock wren populations from not effectively controlling predators (Michelsen-Heath and Gaze 2007; Weston et al. 2018; Rawlence 2019), we believe that repeating the methodology of this study multiple times to increase confidence that mortality is less than 3% should not be the highest research priority for this species. Instead, we suggest that monitoring rock wren population trends using simple transect counts (Monks et al. 2021) would be better for understanding the efficacy of 1080 as an alpine predator management tool over time. For the same investment as this study, conservation managers could monitor population trends at about ten sites, while retaining the ability to ascertain whether aerial 1080 operations delivered under certain environmental conditions result in declines in rock wren populations. Identification of factors that may result in declining rock wren populations would likely occur sooner using transect counts due to the increased sample size of sites, 1080 operations, weather, and their interaction. In summary, we recommend continuing an adaptive management approach to predator control, with population level rock wren monitoring at treatment and non-treatment sites.

Acknowledgments

The ethics application for this work was prepared by Kerry Weston. Thanks also to Graeme Elliott, Martin Genet, Alec Milne, Marian Milne, Richard Stocker, Nina Visker, and Kerry Weston for their assistance in the field. Anja McDonald provided the illustration in Figure 1. Comments from Craig Gillies, Clayson Howell, and two anonymous referees greatly improved this manuscript.

Additional information and declarations

Data and code availability: The data from this article is available from the authors.

Author contributions: TR designed the study. TR and KS led and undertook the field research. TR wrote the manuscript.

Ethics: This work was carried out under DOC AEC approval number AEC 331.

Conflicts of interest: The authors declare no conflicts of interest.

Funding: Funding for this project was provided by the Department of Conservation through the landscape-scale predator control research fund.

References

- Barron DG, Brawn J, Weatherhead PJ 2010. Meta-analysis of transmitter effects on avian behaviour and ecology. *Methods in Ecology and Evolution* 1(2): 180–187.
- Bell EA, Bell BD, Merton DV 2016. The legacy of Big South Cape: rat irruption to rat eradication. *New Zealand Journal of Ecology* 40(2): 212–218.
- Bowen L, Morgan D, Eason C 1995. Persistence of sodium monofluoroacetate (1080) in baits under simulated rainfall. *New Zealand Journal of Agricultural Research* 38(4): 529–531.
- Buller WL 1873. *A history of the birds of New Zealand*. London, Taylor and Francis. 256 p.
- Carpenter JK, Monks A, Innes J, Griffiths J, Anderson D 2023. Immigration drives ship rat population irruptions in marginal high-elevation habitat in response to pulsed resources. *Ecosphere* 14(2): e4424.
- Christie JE, Wilson PR, Taylor RH, Elliott G 2017. How elevation affects ship rat (*Rattus rattus*) capture patterns, Mt Misery, New Zealand. *New Zealand Journal of Ecology* 41(1): 113–119.
- Department of Conservation 2022. Department of Conservation Te Papa Atawhai Annual Report Pūrongo-ā-tau for the year ended 30 June 2022. Wellington, Department of Conservation. 151 p.
- Eason CT, Gooneratne R, Wright GR, Pierce R, Frampton CM 1993. The fate of sodium monofluoroacetate (1080) in water, mammals and invertebrates. *Proceedings of the New Zealand Plant Protection Conference* 46: 297–301.
- Elliott G, Kemp J 2016. Large-scale pest control in New Zealand beech forests. *Ecological Management & Restoration* 17(3): 200–209.
- Fox NC 1977. *The biology of the New Zealand falcon (Falco novaeseelandiae Gemlin 1977)*. Unpublished PhD thesis, University of Canterbury, Christchurch.

- Galbreath R, Brown D 2004. The tale of the lighthouse-keeper's cat: discovery and extinction of the Stephens Island wren (*Traversia lyalli*). *Notornis* 51(4): 193–200.
- Guthrie-Smith H 1936. Sorrows and joys of a New Zealand naturalist. Wellington, AH & AW Reed. 252 p.
- Heath S 1989. The breeding biology of the rock wren, *Xenicus gilviventris*, in the Murchison Mountains, Fiordland National Park, South Island, New Zealand. Unpublished MSc thesis, University of Otago, Dunedin.
- Holdaway RN 1999. Introduced predators and avifaunal extinction in New Zealand. In: Extinctions in near time: causes, contexts, and consequences. Boston, MA, Springer US. Pp. 189–238.
- Kemp JR, Mosen CC, Elliott GP, Hunter CM 2018. Effects of the aerial application of 1080 to control pest mammals on kea reproductive success. *New Zealand Journal of Ecology* 42(2): 158–168.
- Kemp JR, Mosen CC, Elliott GP, Hunter CM, van Klink P 2019. Kea survival during aerial poisoning for rat and possum control. *New Zealand Journal of Ecology* 43(1): 3351.
- McNab BK, Weston KA 2020. Does the New Zealand rockwren (*Xenicus gilviventris*) hibernate? *Journal of Experimental Biology* 223(9): 121–126.
- Michelsen-Heath S, Gaze P 2007. Changes in abundance and distribution of the rock wren (*Xenicus gilviventris*) in the South Island, New Zealand. *Notornis* 54(2): 71–78.
- Mitchell KJ, Wood JR, Llamas B, McLenachan PA, Kardailsky O, Scofield RP, Worthy TH, Cooper A 2016. Ancient mitochondrial genomes clarify the evolutionary history of New Zealand's enigmatic acanthisittid wrens. *Molecular Phylogenetics and Evolution* 102: 295–304.
- Monks JM, O'Donnell CF, Greene TC, Weston KA 2021. Evaluation of counting methods for monitoring populations of a cryptic alpine passerine, the rock wren (Passeriformes, Acanthisittidae, *Xenicus gilviventris*). *Plos one* 16(3): e0247873.
- Nugent G, Morriss GA, Warburton B 2019. Attempting local elimination of possums (and rats) using dual aerial 1080 baiting. *New Zealand Journal of Ecology* 43(2): 3373.
- O'Donnell CF, Weston KA, Monks JM 2017. Impacts of introduced mammalian predators on New Zealand's alpine fauna. *New Zealand Journal of Ecology* 41(1): 1–22.
- Powlesland R, Knegtman J, Marshall I 1999. Costs and benefits of aerial 1080 possum control operations using carrot baits to North Island robins (*Petroica australis longipes*), Pureora Forest Park. *New Zealand Journal of Ecology* 23(2): 149–159.
- Powlesland R, Knegtman J, Styche A 2000. Mortality of North Island tomtits (*Petroica macrocephala toitoi*) caused by aerial 1080 possum control operations, 1997–98, Pureora Forest Park. *New Zealand Journal of Ecology* 24(2): 161–168.
- Rappole JH, Tipton AR 1991. New harness design for attachment of radio transmitters to small passerines. *Journal of Field Ornithology* 62(3): 335–337.
- Rawlence TE 2019. The efficacy of aerial 1080 poison applied on a landscape scale to control alpine predators and the reproductive response of rock wren (*Xenicus gilviventris*). Unpublished MSc thesis, University of Otago, Dunedin.
- Robertson HA, Baird KA, Elliott GP, Hitchmough RA, McArthur NJ, Makan TD, Miskelly CM, O'Donnell CFJ, Sagar PM, Scofield RP, Taylor GA, Michel P 2021. Conservation status of New Zealand birds 2021. *New Zealand Threat Classification Series* 36. Department of Conservation, Wellington. 43 p.
- Russell JC, Innes JG, Brown PH, Byrom AE 2015. Predator-free New Zealand: conservation country. *BioScience* 65(5): 520–525.
- Stocker R, Petyt C, Milne M 2006. A five-year study of rock wren in Henderson Basin, Kahurangi National Park. Department of Conservation, Nelson/Marlborough Conservancy. 23 p.
- Streby HM, McAllister TL, Peterson SM, Kramer GR, Lehman JA, Andersen DE 2015. Minimizing marker mass and handling time when attaching radio-transmitters and geolocators to small songbirds. *The Condor: Ornithological Applications* 117(2): 249–255.
- Veltman CJ, Westbrooke IM 2011. Forest bird mortality and baiting practices in New Zealand aerial 1080 operations from 1986 to 2009. *New Zealand Journal of Ecology* 35(1): 21–29.
- Veltman CJ, Westbrooke IM, Powlesland RG, Greene TC 2014. A principles-based decision tree for future investigations of native New Zealand birds during aerial 1080 operations. *New Zealand Journal of Ecology* 38(1): 103–109.
- Weston K, O'Donnell C, van Dam-Bates P, Monks J 2018. Control of invasive predators improves breeding success of an endangered alpine passerine. *Ibis* 160: 892–899.
- Wright J 2011. Evaluating the use of 1080: predators, poisons and silent forests. Wellington, New Zealand Parliamentary Commissioner for the Environment. 85 p.

Received: 7 November 2023; accepted: 19 April 2024

Editorial board member: Warwick Allen